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Habitat type and environmental conditions influence the age and growth of a temperate marine damselfish

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Life history parameters for fishes have generally been applied to species across their entire range, however, different ecological and environmental conditions and processes (e.g., sea surface temperature, habitat, primary productivity, fishing mortality, resource availability) influence life history patterns at smaller spatial scales. By focusing on a historically protected species, we determined how environmental and ecological factors shape patterns in growth and longevity, without the impact of fishing. The Garibaldi (*Hypsypops rubicundus*) is a territorial marine damselfish native to the shallow rocky reefs of southern California. Garibaldi were collected from five mainland locations and five Channel Islands throughout the Southern California Bight. Paired natural reef and artificial reef habitats (i.e., breakwaters and jetties) in each mainland location were sampled. Otolith-based ageing and biological data from these populations were used to assess how age and growth vary by location, reef type (natural/artificial), island or mainland, mean annual sea surface temperature, and/or sex. The annual formation of growth increments in otoliths was validated *in-situ* using tetracycline mark-recapture methods to confirm that increments are formed annually. Garibaldi grew significantly larger on artificial reefs than on natural reefs but tended to live longer on paired natural reefs. Regionally, growth and longevity followed mean annual sea surface temperature gradients, consistent with Bergmann's rule. Garibaldi exhibited clear sexual size dimorphism; males grew larger, which is uncharacteristic for both damselfish and other marine fishes from the southern California. The Garibaldi collected for this study had a maximum age of 57 years, which makes this the longest lived damselfish species by two decades.

KEYWORDS

artificial reefs, southern California, fish life history, growth patterns, spatial variation, Garibaldi, *Hypsypops rubicundus*

Introduction

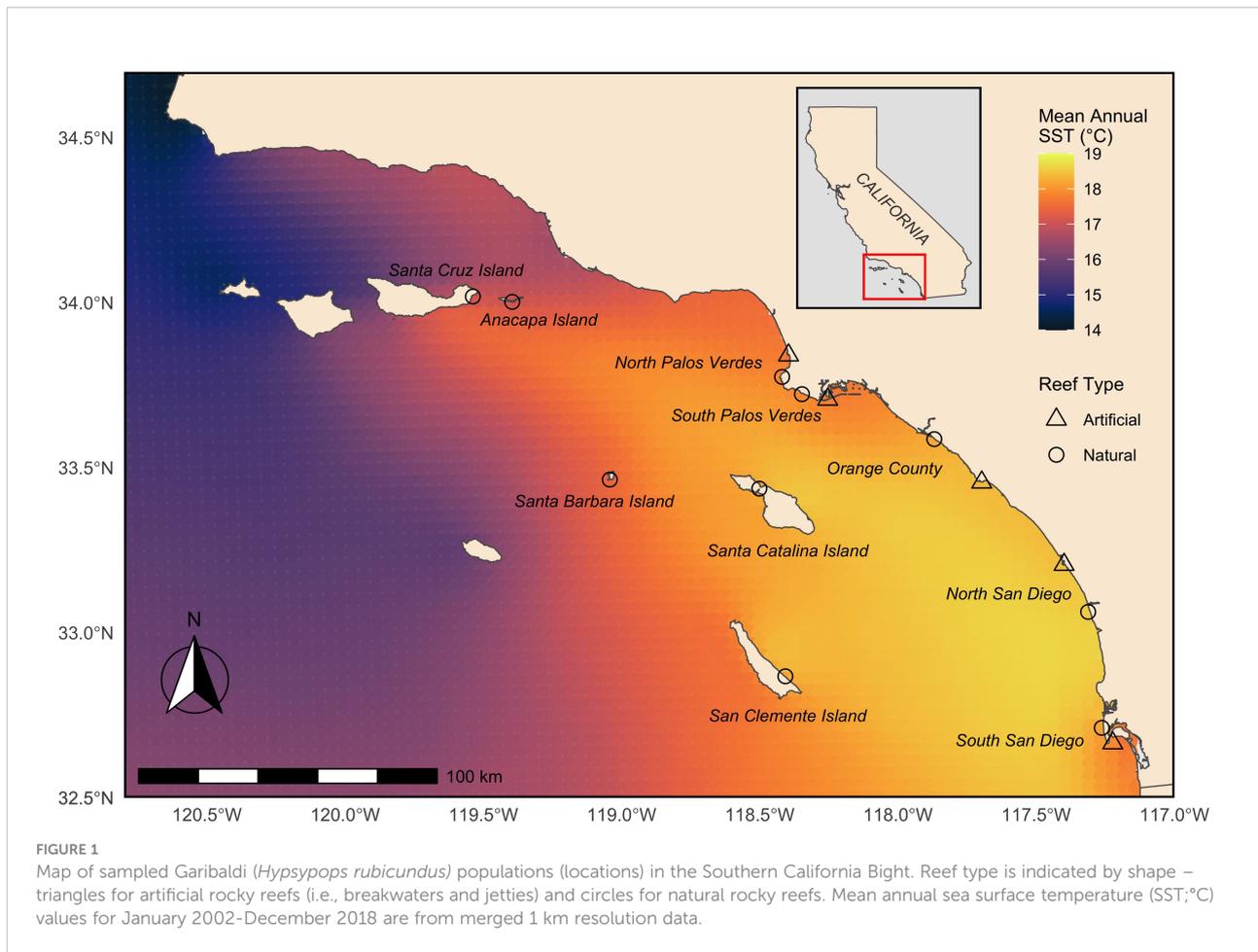
Marine fish life history strategies (i.e., the allocation of limited resources to the competing demands of growth, reproduction, and survival) and demographic parameters can provide insight into population and community-level processes and ecological patterns. They are typically assumed to be consistent for a species across its entire range (King and McFarlane, 2003; Fonseca and Cabral, 2007). However, several recent studies concluded that different ecological and environmental conditions result in intraspecific spatial variation in fish life history patterns. This variation has been attributed to multiple factors including: differences in mean annual, inter-annual, or range in sea surface temperature (Trip et al., 2014; Trip et al., 2016; Tanner et al., 2019); habitat (DeMartini et al., 1989; DeMartini et al., 1994; Love et al., 2007; Granneman and Steele, 2014); primary productivity (Ruttenberg et al., 2005; Tanner et al., 2019); fishing mortality (Robertson et al., 2005; Hamilton et al., 2007; Hsieh et al., 2009; Hamilton et al., 2011; Taylor, 2013; Taylor et al., 2015); differing environmental conditions during larval/juvenile growth (Atkinson, 1994; Atkinson, 1996; Fonseca and Cabral, 2007; Williams et al., 2007); food resource availability (Hjelm et al., 2000); and density-dependent interactions (Figueira et al., 2007). Understanding the complexities of species-specific life histories for marine fishes is essential for informing ecosystem-based fisheries management (Jennings and Dulvy, 2008).

The Southern California Bight (SCB) spans 3.5° of latitude of the ocean from Point Conception, CA in the north to the United States-Mexico border in the south. The SCB encompasses 1,198 km of coastline that includes the mainland and an eight-island archipelago, known as the California Channel Islands (Pondella et al., 2015b). Roughly 26% of the nearshore (<30 m depth) habitat is composed of rocky reefs with the remainder dominated by soft substrata, exhibiting fragmented distributions of rocky reef habitat. The SCB has a unique surface current circulation pattern where the year-round, dominant California Current brings cool water to the region from the north. South of Point Conception, the flow of the California Current is disrupted by the Channel Islands and branches shoreward and poleward where it forms the Southern California Countercurrent, drawing warmer water from the south and creating a large eddy that circulates counterclockwise throughout the Bight (Hickey, 1992). This current pattern, when coupled with the complex bathymetry, results in strong abiotic gradients such as sea surface temperature (Figure 1) and diverse habitats that support various assemblages of fish fauna. This provides a prime opportunity to study how variability in environment and habitat type shape life history patterns (Engle, 1993; Allen and Pondella, 2006; Pondella et al., 2015a; Pondella et al., 2015b). Prior studies of spatial differences in growth patterns and

demographic parameters for fishes in the SCB have concentrated on commercially and recreationally exploited species (Love et al., 1990; Love et al., 2007; Granneman, 2011; Hamilton et al., 2011; Granneman and Steele, 2014). Size-selective fishing mortality on larger, older members of a population can influence fish life history patterns including rapid growth to asymptotic size, early maturity, reduced reproductive output, skewed sex ratios, and age- and size-truncation of exploited populations, which can have evolutionary consequences (Robertson et al., 2005; Hamilton et al., 2007; Hsieh et al., 2009; Caselle et al., 2011; Hamilton et al., 2011; Taylor, 2013; Audzijonyte et al., 2016).

In addition to natural rocky reefs, artificial or anthropogenic structures also provide habitat for reef-associated fishes in the SCB. Over the last 50 years, the population of southern California has nearly doubled and is one of the most densely populated areas in the United States, currently supporting roughly 24 million people. In addition, nearly 25% of the entire nation's coastal population lives in the four counties bordering the shoreline of the SCB (Schiff et al., 2000; Wilson and Fischetti, 2010). The artificial structures in the SCB were primarily built as coastal infrastructure (e.g., breakwaters and jetties, oil and gas platforms, and outfall and discharge pipes) to support this urbanization. In addition, rapid population growth has put stress on marine resources through the discharge of pollutants, loss of habitat, and overfishing. In 1958, California Department of Fish and Game (currently California Department of Fish and Wildlife) began the Nearshore Sportfish Habitat Enhancement Program, which focused on restoring and enhancing fish habitat through the implementation of purpose-built artificial reefs in southern California (Lewis and McKee, 1989). The function of artificial structures, both infrastructure and purpose-built, as fish habitat is well-studied within the Bight (Ambrose and Swarbrick, 1989; DeMartini et al., 1989; DeMartini et al., 1994; Pondella and Stephens, 1994; Stephens et al., 1994; Pondella et al., 2002; Stephens and Pondella, 2002; Claisse et al., 2014; Granneman and Steele, 2015; Claisse et al., 2019; Love et al., 2019; Barilotti et al., 2020; Burns et al., 2020), however relatively few studies have compared the ecological performance of fish on artificial reefs to natural reefs (Love et al., 2007; Granneman and Steele, 2014).

The Garibaldi (*Hypsypops rubicundus*) is a territorial marine damselfish endemic to the shallow reefs of southern California and Baja California, Mexico. Out of 385 species of damselfish (Family: Pomacentridae), Garibaldi is the largest, reaching up to 35 cm in length (Limbaugh, 1964; Clarke, 1970; Allen, 1991; Nelson et al., 2016). Their bright orange adult coloration, with spots of iridescent blue as juveniles, makes them the most conspicuous reef fish in southern California (Kritzler et al., 1950). Once an intense target of collection for the aquarium trade, as the State Marine Fish of California, this species has been fully protected in the United States since 1995. Although



predation on young-of-the-year Garibaldi by Kelp Bass (*Paralabrax clathratus*) (Cook, 2011), adult Garibaldi have very few known natural predators (Kritzler et al., 1950; Clarke, 1970) and are currently unfished. Unlike other damselfish where males guard nest sites only during the mating season, both male and female adult Garibaldi typically defend year-round territories of approximately 3–10 m², demonstrating high site fidelity (Limbaugh, 1964; Clarke, 1970; Foster, 1972; Sikkel, 1988; Sikkel, 1989; Alcalay and Sikkel, 1994). A territory includes a shelter hole, foraging area, and for adult males, a nest site comprised of perennial patches of red algae cultivated by the male. Females are allowed into the territory to deposit eggs into the nest during spawning season (April through September), the timing of which varies within and among populations (Clarke, 1970; Clarke, 1971; Sikkel, 1988). The eggs are guarded by the male until young hatch, and disperse into the plankton (Clarke, 1970). In early summer to early fall, juveniles settle on shallow rocky reefs on exposed or semi-protected coasts (Limbaugh, 1964; Clarke, 1970; Alcalay and Sikkel, 1994). Vagility is only through the larval stage making this taxon optimal for describing site-specific growth characteristics. Previous life history studies on Garibaldi are

limited primarily to the San Diego, CA area and/or prior to 1995, when their protected status was implemented. Clarke (1970) reported that Garibaldi achieve adult coloration at around 5 years, sexual maturity at 6 years, and a maximum age of 17 years. However, Clarke's study used scale-based ageing, which can drastically underestimate the ages of fish (Francis and Francis, 1992; Beamish and McFarlane, 1995; Campana, 2001), and was conducted prior to the implementation of protected status. During the period from when that study was conducted in 1965 to 2010, Garibaldi densities more than doubled at Clarke's study sites (Cook, 2011), indicating that the status of Garibaldi populations have changed considerably over time. By focusing on a historically protected species with high site fidelity, like the Garibaldi, we can determine how habitat type, environmental, and ecological factors shape patterns in growth and longevity at distinct locations within the SCB.

For this study, otolith-based ageing was used to assess how life history patterns vary by reef type (natural/artificial), mean annual sea surface temperature, sex, island or mainland geographies, and/or location for Garibaldi populations throughout the SCB. The periodicity of increment formation in otoliths was assessed using *in-situ* tetracycline mark-recapture

methods to confirm that increments form annually rather than conforming to some other periodicity (e.g., seasonality, lunar cycles) (Campana, 2001). Accounting for patterns in variation of fish life history observed in previous studies, the following predictions were made:

1. Garibaldi from artificial reef habitats will grow larger and faster than Garibaldi from natural reef habitats. Within the SCB, juvenile (young-of-the-year) Blue Rockfish (*Sebastes mystinus*) were found to have faster daily growth rates at oil platforms compared to nearby natural reefs (Love et al., 2007). Garibaldi were collected from breakwaters because they were consistent in habitat structure and presence in each mainland location. Breakwaters and jetties are built to protect coastal infrastructure and are made up of large (1-2 m) quarry rock boulders, providing a complex and ideal fish habitat with several open spaces to provide refuge and for Garibaldi to maintain territories (Stephens et al., 1994).

2. Garibaldi will be larger, slower-growing, and longer-lived at cooler locations, and smaller, faster-growing, and shorter-lived at warmer locations. Ectotherms tend to be larger and live longer in colder environments due to a trade-off between metabolism and longevity (Atkinson and Sibly, 1997; Mangel and Abrahams, 2001). In addition, fish populations in colder waters often have shorter reproductive seasons and therefore may allocate less of their annual energy budgets to reproduction, which results in larger and older fish in these populations (Ruttenberg et al., 2005).

3. Male Garibaldi will grow larger than female Garibaldi. Although previous Garibaldi life history studies found no apparent differences between growth rates for males and females, they did find that males lived longer and grew larger (Clarke, 1970). In addition, Sikkil (1988; 1989) determined that male size was positively correlated with female nest preference and posited that male size may be correlated with nest quality. Further, large males may be more successful in the defense of their nesting sites.

Methods

Study area

Garibaldi were collected from ten locations in the SCB (Figure 1). Locations were selected based on two criteria: they supported at minimum 100 Garibaldi per hectare (Pondella et al., 2015a); and were ecologically and environmentally distinct and geographically distant (Engle, 1993; Pondella et al., 2015b). Five locations were in the California Channel Islands of Santa Cruz Island, Anacapa Island, Santa Barbara Island, Santa Catalina Island, and San Clemente Island and the other five locations were on the mainland coast in Northern and Southern Palos Verdes, Orange County, and Northern and Southern San Diego. For locations on the mainland, paired natural rocky reefs and artificial or human-built rocky reefs

(e.g., breakwaters and jetties at harbors nearby or adjacent to natural reef sites) were sampled. This investigated the influence of differences in habitat type under similar environmental conditions.

Fish collections

From 2013–2019, 1,111 Garibaldi were collected from the ten study locations (Figure 1) by scientific divers using pole spears. Divers attempted to collect a representative size sample of the fish present at each location. Upon collection, fish were transported to the laboratory where they were weighed to the nearest gram and measured head (HL), standard (SL), fork (FL), and total lengths (TL) to the nearest millimeter. The digestive tract and gonads were removed, weighed, and preserved for future studies. Sex was determined macroscopically, if possible. Macroscopic sex determination was difficult outside of the spawning season, especially in smaller male individuals, therefore sex was determined through histological preparations when necessary. Sagittal otoliths were extracted and stored dry. This study was reviewed and approved by the Institutional Care and Use Committees at both Cal Poly Pomona and Occidental College, and fish were collected under California Department of Fish and Wildlife Scientific Collecting Permits.

Otolith processing and age determination

Sagittal otoliths were used for ageing. Annual periodicity of sagittal otolith increment formation was validated using standard methods (Cappo et al., 2000) including *in-situ* tetracycline mark-recapture at Flat Rock in North Palos Verdes (Section 1 in Supplementary Material). Each whole sagittal otolith was weighed to the nearest 0.0001 g. One otolith from each fish was mounted on a glass slide using thermoplastic resin and sectioned using a 1200 grit diamond flat lap on a grinding wheel following Taylor and McIlwain (2010) to produce a thin transverse section. Next, otoliths were covered in resin and heated on a hotplate to 260 °C, to improve readability. This increased the contrast between alternating translucent and opaque annual zones producing results similar to the “break and burn method” (Forsberg, 2001). When the transverse sections were viewed under a stereomicroscope with reflected light, each displays a central, dense, opaque region (pre-settlement) and alternating translucent and opaque zones (annuli) (Figure S2 in Supplementary Material). Annuli were counted along a path from the center towards the tip of the inner face next to the sulcus, where annual growth increments are clearly defined (García-Mederos et al., 2016). Readings were done by a single reader (C. M. Williams) a minimum of two times. The final age (in years) of an individual was determined when two or more counts agree (Choat and Axe, 1996; Claisse

et al., 2009). If at least two counts did not agree after three readings, the otolith was excluded from age-based analyses (1.1% of otoliths).

Age and growth analyses

All age and growth analyses were performed in R (R Core Development Team, 2021) and are reported in relation to fish total length (TL; mm). Growth patterns were evaluated by fitting the von Bertalanffy growth function (VBGF) using maximum likelihood using the ‘mle2’ function in the *bbmle* package (Bolker and R Core Development Team, 2020), which models the mean TL (L_t , in mm) at age (t , in years) following:

$$L_t = L_\infty [1 - e^{-k(t-t_0)}]$$

where L_∞ is mean asymptotic length, k describes the rate at which L_t reaches L_∞ (with units of yr^{-1}), and t_0 is the theoretical age at 0 mm in length. Since fish often demonstrate rapid initial growth, it is important to include small, young fish to obtain accurate estimations of mean size-at-age for early life stages when modeling growth (Kritzer et al., 2001; Berumen, 2005; Claisse, 2009). Therefore, for immature individuals where sex was unidentifiable *via* macroscopic identification and histological preparation, sex was assigned randomly following a 1:1 sex ratio. Additionally, because there were few individuals of age 0-3 within each distinct grouping, t_0 was fixed to a value of -2.1 years which was determined by fitting the von Bertalanffy growth function to the global population.

The longevity of fishes is typically assessed using three methods: (1) the maximum age, or oldest individual in the population (T_{\max}) (Beverton and Holt, 1959), (2) as the mean age of the oldest 25% of fish collected ($T_{\max\ 25\%}$) (Choat and Robertson, 2002), and (3) as the mean age of the oldest 10% of fish collected ($T_{\max\ 10\%}$) (Trip et al., 2008). All three metrics were calculated for males and females for each location and reef type, and for the global population.

Environmental data

Measures of environmental variables were derived from remotely sensed data. Mean annual sea surface temperature (SST; °C) values were derived from merged MODIS 1 km resolution data from MODIS-Aqua and MODIS-Terra sensors composited over each month from January 2002- December 2018. This period was used to best represent the average conditions that fish experienced during their lifetime, with the data available from MODIS, which spans 2002-present. Surface chlorophyll-a (Chl-a; g C m^{-3}) values for each location were derived using satellite data from MODIS-Aqua, MODIS-Terra, VIIRS-NPP, SeaWiFS, MERIS and OLCI sensors composited

over each month during the same time period. Data were merged using simple pixel-wise averaging of valid pixel values to increase coverage and composited by the California Current Ecosystem Long-Term Ecological Research program based at Scripps Institution of Oceanography. Full-resolution 1 km data sets were downloaded in HDF5 format, clipped to the study area then converted to a data frame table format for plotting specific points using R (R Core Development Team, 2021). Values for mean annual SST and Chl-a were calculated by averaging the annual means. Because nearshore values for both Chl-a and SST often contain erroneous values due to some cells containing a portion of land, these data were then imported into ArcGIS, the cells that overlapped with land were deleted, and then the layers were converted to a 4 km raster using the Point to Raster tool (ESRI ArcMap 10.8.1 Spatial Analyst) and assigned a value using the average values of all the points within the cell. Cell values for SST and Chl-a for each study location were then extracted using the Extract Values to Points tool.

Age and growth model selection and confidence interval estimation

The level of evidence for effects of sex, reef type, island/mainland, and location on growth patterns was investigated using model selection. We used the corrected Akaike Information Criterion (AICc), which adds a correction for bias due to small sample size, effectively penalizing complex models with small data sets (Burnham and Anderson, 2002; Anderson, 2008). Akaike weights (w_i) were calculated to assess the relative likelihood of each model in the set and were interpreted as the weight of evidence in favor of the model. Parameter estimates are reported for the highest ranked model, as indicated by w_b , and 95% confidence intervals were produced using random bootstrap resampling with replacement (5,000 bootstraps) using the ‘bootstraps’ function in the *rsample* package (Silge et al., 2021) and mapped to the grouped data using the ‘map’ function in the *purrr* package (Henry and Wickham, 2020). Sampling with replacement assumes that the variation in the sampled group is representative of the population as a whole, and any additional fish collected would fall within the bounds of the sampled group (Kritzer et al., 2001). For models fit to data subsets with small sample sizes where bootstrap confidence intervals produced unrealistic values, the 95% confidence intervals from the likelihood profile are reported.

In total, 1,027 fish were used in the final data set for model selection. The sampled population from Anacapa Island was small ($n = 55$) so it was grouped with the sampled population from Santa Cruz Island ($n = 101$) for geographic analyses. Anacapa Island and Santa Cruz Island sites were close to each other and had similar environmental conditions and habitat types (Table 1 and Figure 1). Surface Chl-a was strongly

TABLE 1 Garibaldi (*Hypsypops rubicundus*) collection locations, environmental and geographic information including mean annual sea surface temperature (SST; °C) and mean annual Chlorophyll-a (Chl-a, g C m⁻³) from January 2002- December 2018.

Location	Reef type	Latitude	Longitude	Mean annual SST (°C)	Mean annual Chl-a (g C m ⁻³)	Dates sampled
Santa Cruz Island	Natural	34.01926	-119.53821	17.05	1.233	Aug 2015*, Aug 2017
Anacapa Island	Natural	34.00289	-119.39580	17.15	0.992	Aug 2017
Santa Barbara Island	Natural	33.46439	-119.04280	17.51	0.654	Jan 2019
Santa Catalina Island	Natural	33.43702	-118.50201	18.24	0.607	Aug 2013**, Jul 2014
North Palos Verdes	Natural	33.77567	-118.41960	17.63	2.405	Aug 2013, Jun/Jul 2014, Dec 2015, Nov 2019
King Harbor Breakwater	Artificial	33.84133	-118.39630			Aug/Sept 2013, Mar 2014, Jun 2015, Jan 2016, Mar 2017
South Palos Verdes	Natural	33.72333	-118.34780	17.87	2.166	Oct 2013, Dec 2015
Port of Los Angeles Breakwater	Artificial	33.70726	-118.25460			Jun 2016
Orange County	Natural	33.58706	-117.86900	18.30	1.820	Sept 2018
Dana Point Harbor Breakwater	Artificial	33.45614	-117.69730			Aug 2015, Jun 2018
San Clemente Island	Natural	32.86839	-118.40780	18.25	0.434	Feb 2019
North San Diego	Natural	33.06349	-117.31230	18.48	2.133	May 2018
Oceanside Harbor Breakwater	Artificial	33.20677	-117.39940			May 2018
South San Diego	Natural	32.71210	-117.26300	17.81	2.947	Jun 2018
Zuniga Jetty	Artificial	32.66584	-117.22320			Jun 2018

*Collected by Ben Grime, Kelp Forest Monitoring Program, U.S. National Park Service.

**Collected by Dr. Scott Hamilton, Moss Landing Marine Laboratories, San José State University.

All collections were made by the Vantuna Research Group unless otherwise noted.

correlated with the island/mainland categorization, likely due to localized upwelling events common at island sites (Kilpatrick et al., 2018) and terrestrial runoff at mainland sites (Schiff et al., 2000), so only island/mainland was used as a variable with the understanding that mainland sites often have significantly higher primary productivity than island sites.

Results

There is strong evidence that Garibaldi age and growth patterns vary by sex, reef type, mean annual sea surface temperature, and location within the SCB. The highest ranked model, where model parameters were fitted separately for each sex at each reef type within each location (Table 2; Figure S5 in Supplementary Material), was also the most complex model and received 100% of total Akaike weight (w_i) (Table S4 in Supplementary Material).

Garibaldi demonstrated clear sexual dimorphism in body length with males exhibiting greater mean asymptotic total length (L_∞) than females at each location for each reef type, with only one exception, the artificial reef sampled in North San Diego, Oceanside Harbor Breakwater (Table 2; Figures 2A, S5 in Supplementary Material). In most instances, male-female pairwise L_∞ 95% confidence intervals did not overlap (Table 2). When data from all reef types and locations were pooled, mean asymptotic length (L_∞) was significantly higher for males (271 mm, 95% CI: 267 to 274 mm) than females (247 mm, 95% CI: 243 to 250 mm), with adult males being 24 mm larger than

females on average (Table S5; Figure S8 in Supplementary Material). The greatest difference in L_∞ between sexes was observed at the natural reef site in Orange County, where L_∞ for males was 42 mm larger than for females. In contrast, the L_∞ values between sexes were most similar at the artificial reef in South Palos Verdes, the Port of Los Angeles Breakwater, which differed by only 2 mm (Table 2).

Garibaldi grew significantly larger on artificial reefs than on paired natural reefs across the mainland locations, with females being 26 mm larger (95% CI: 20 to 32 mm) and males 15 mm larger (95% CI: 10 to 20 mm) on average, when data from all sites were pooled (pairwise differences in L_∞ ; Figure 2B). For the smaller location-specific samples, there were significant pairwise differences (95% CIs did not include 0) for females at all but one location, (mean pairwise difference range: 18 to 36 mm), and males exhibited positive, albeit smaller and not significant pairwise differences at all but one mainland location (mean pairwise difference range: -2 to 24 mm). Female L_∞ values from all but one of the artificial reefs were also higher than any natural reef populations including fish from Anacapa and Santa Cruz Islands, which was the northernmost and coolest location sampled (Figure 3). Additionally, the L_∞ of females from the South Palos Verdes artificial reef (i.e., the Los Angeles Harbor breakwater) was still higher than those from almost all natural reefs. In general, males exhibited a similar, but less distinct pattern, interrupted by males from the artificial reef in South Palos Verdes exhibiting relatively low L_∞ and males from natural reef in Orange County and South San Diego exhibiting relatively high L_∞ values (Figure 3).

TABLE 2 Garibaldi (*Hypsypops rubicundus*) von Bertalanffy growth function parameters (bold) for the highest ranked (and most complex) model with bootstrapped 95% confidence intervals.

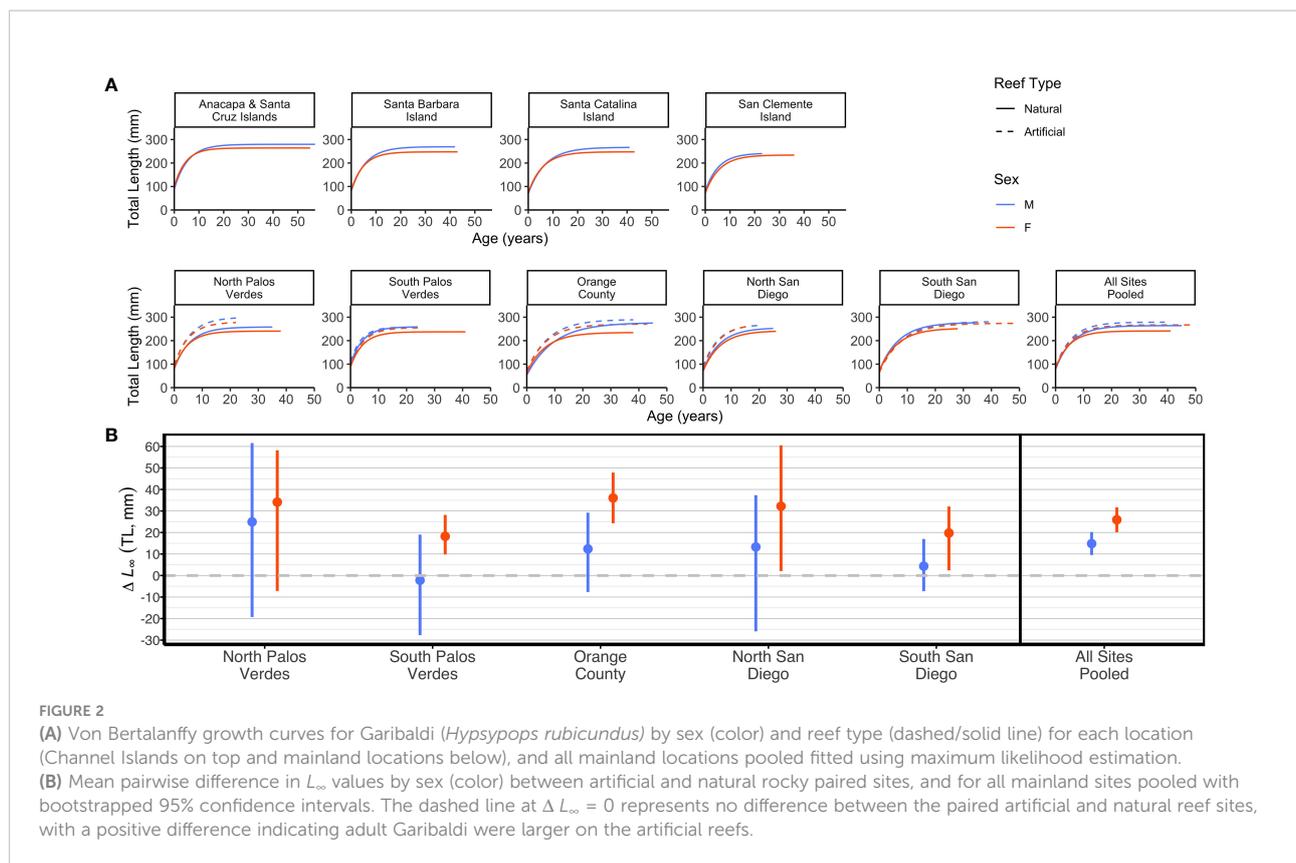
Population					Parameters and 95% Confidence Intervals					
Location	Reef type	Sex	n	Size (TL mm)	L_{∞} (mm)	95% CI	k (yr ⁻¹)	95% CI	σ	95% CI
Anacapa & Santa Cruz Islands	Natural	M	65	98 - 307	280	274 - 286	0.18	0.17 - 0.20	24	19 - 28
		F	87	18 - 290	264	259 - 269	0.22	0.20 - 0.24	26	21 - 30
Santa Barbara Island	Natural	M	32	139 - 283	269	260 - 277	0.17	0.16 - 0.19	14	10 - 18
		F	37	122 - 261	248	235 - 257	0.20	0.18 - 0.24	23	18 - 26
Santa Catalina Island	Natural	M	70	29 - 285	266	262 - 273	0.15	0.13 - 0.16	18	15 - 20
		F	77	29 - 278	247	243 - 253	0.17	0.15 - 0.19	17	14 - 19
North Palos Verdes	Natural	M	34	119 - 273	258	244 - 270	0.18	0.15 - 0.22	19	14 - 23
		F	46	107 - 284	241	236 - 248	0.20	0.18 - 0.23	17	11 - 21
	Artificial	M	35	116 - 298	301	243 - 315	0.18	0.15 - 0.21 *	20	14 - 57
South Palos Verdes	Natural	M	16	180 - 270	259	241 - 276	0.23	0.20 - 0.28 *	12	7 - 23
		F	30	77 - 250	238	234 - 242	0.23	0.19 - 0.27	14	9 - 17
	Artificial	M	32	98 - 278	258	233 - 268	0.27	0.23 - 0.31 *	15	11 - 31
Orange County	Natural	M	26	127 - 286	277	265 - 297	0.10	0.08 - 0.12	15	11 - 19
		F	31	118 - 253	235	225 - 246	0.15	0.13 - 0.18	15	11 - 18
	Artificial	M	41	120 - 296	290	282 - 301	0.13	0.11 - 0.15	18	14 - 21
San Clemente Island	Natural	M	33	148 - 247	242	233 - 254	0.19	0.15 - 0.24	19	12 - 24
		F	38	126 - 241	234	229 - 239	0.17	0.15 - 0.20	13	9 - 17
	Artificial	M	32	123 - 276	269	231 - 287	0.19	0.14 - 0.25 *	28	18 - 37
North San Diego	Natural	M	19	130 - 254	255	240 - 266	0.16	0.13 - 0.21	21	11 - 29
		F	40	118 - 240	242	237 - 249	0.16	0.15 - 0.18	11	9 - 14
	Artificial	M	32	123 - 276	269	231 - 287	0.19	0.14 - 0.25 *	28	18 - 37
South San Diego	Natural	M	25	132 - 286	278	267 - 287	0.14	0.12 - 0.16	15	9 - 19
		F	23	122 - 243	253	245 - 271	0.15	0.13 - 0.18	13	9 - 15
	Artificial	M	45	135 - 287	282	274 - 286	0.12	0.11 - 0.15	14	11 - 17
		F	22	134 - 288	274	259 - 269	0.13	0.11 - 0.14	12	8 - 15

L_{∞} is mean asymptotic total length in mm, k describes the yearly rate at which mean TL approaches L_{∞} , σ is the standard deviation of total length at a given age. t_0 was not reported because it was fixed to a value of -2.1 years which was determined by fitting the von Bertalanffy growth function to the global population. Bootstrap 95% confidence intervals were reported except in cases where they were unrealistic (noted with an asterisk *), in which case the likelihood profile 95% CIs are reported.

The negative relationship between mean asymptotic total length (L_{∞}) and mean annual SST was much stronger for natural reefs (male $R^2 = 0.22$, slope = -12.9 mm/°C, 95% CI: -34.6 to 8.8 mm/°C; female $R^2 = 0.49$, slope = -14.8 mm/°C, 95% CI: -28.4 to -1.3 mm/°C) than at artificial reefs (male $R^2 = 0.09$, slope = -14.2 mm/°C, 95% CI: -97.1 to 68.8 mm/°C; female $R^2 = < 0.01$, slope = -1.55 mm/°C, 95% CI: -47.9 to 44.9 mm/°C) for both sexes (Figure 3). Among the Channel Islands, growth patterns followed both a latitudinal and SST gradient (Figure 3). Garibaldi from Anacapa and Santa Cruz Islands, the northernmost and coolest location, had the largest L_{∞} for both sexes at any island (male $L_{\infty} = 280$ mm, female $L_{\infty} = 264$ mm) whereas at the southernmost and warmest channel island, San Clemente Island, L_{∞} was the smallest at any island for both sexes (male $L_{\infty} = 242$ mm, female $L_{\infty} = 234$ mm). Mainland sites spanned a smaller range in both latitude and mean annual SST and the patterns were less distinct. The largest L_{∞} for Garibaldi from

mainland natural reefs was in South San Diego (Point Loma) for both males and females (male $L_{\infty} = 278$ mm, female $L_{\infty} = 253$ mm). The main difference was for males from natural reefs in Orange County, which were similarly large ($L_{\infty} = 277$ mm), while the females were the smallest ($L_{\infty} = 235$ mm). Patterns in the growth coefficient (k) are generally more difficult to interpret than those for L_{∞} because they are dependent on L_{∞} in the von Bertalanffy growth model. However, when k and L_{∞} are plotted together, there was also a positive relationship with temperature for both males and females at natural reefs and a less distinct pattern for artificial reefs (Figure S7 in Supplementary Material).

Longevity (T_{\max}) exhibited a similar pattern with mean annual SST as observed for L_{∞} (Figure 4). Among natural reefs, maximum age (T_{\max}) was negatively correlated with mean annual SST (male $R^2 = 0.36$ slope = -14.9 yrs/°C, 95% CI: -32.8 to 3.0 yrs/°C; female $R^2 = 0.46$, slope = -12.9 yrs/°C, 95% CI: -25.3 to -0.4 yrs/°C). This



supported the hypothesis that fish live longer in colder locations, while there was no relationship for fish from artificial reefs (male $R^2 < 0.01$, slope = 1.31 yrs/ $^{\circ}$ C, 95% CI: -51.5 to 54.1 yrs/ $^{\circ}$ C; female $R^2 = 0.01$, slope = -4.5 yrs/ $^{\circ}$ C, 95% CI: -76.8 to 67.8 yrs/ $^{\circ}$ C) (Figure 4). The maximum ages of Garibaldi in this study were 57 years for males and 55 years for females (Figures 2, 4; Figures S5, S6, and Table S5 in Supplementary Material). Both fish were collected from Santa Cruz Island, the northernmost location with the coldest mean annual SST. Within locations, fish lived longer on natural reefs compared with their paired artificial reefs, except for the South San Diego population (Table S4 in Supplementary Material). There was no apparent difference in longevity between sexes (Figures S5, S6, and Table S6 in Supplementary Material), although the relationship between mean annual SST and maximum age was stronger for females than for males (Figure 4).

Discussion

Spatial and habitat variability in the life history of Garibaldi is apparent at the relatively small spatial scale of this study with growth patterns differing by reef type, location, and sex. Garibaldi exhibited clear sexual size dimorphism; males grew larger than females in all populations. Garibaldi grew significantly larger on artificial quarry rock reefs (i.e.,

breakwaters and jetties) compared to natural rocky reefs, but tended to live longer on natural reefs. Regionally, growth and longevity patterns followed gradients in mean annual sea surface temperature, consistent with the hypothesis that Garibaldi are larger and live longer in cooler locations than in warmer locations. Sex-specific average adult asymptotic size (L_{∞}) was greater for most artificial reef sites than the coldest natural reef sites, indicating, at least across the temperature range sampled, that habitat can have a greater effect than environmental conditions on this life history parameter, particularly for females. The Garibaldi collected for this study had a maximum age of 57 years, which makes them the longest lived of any species of damselfish by two decades (Tzioumis and Kingsford, 1999). Annual otolith increment formation was validated using the *in-situ* tetracycline mark-recapture method which is an ideal technique for age validation for species with high site fidelity (Campana, 2001). This technique is uncommon for fish life history studies conducted in temperate marine systems.

Sexual dimorphism in size was observed within all populations of Garibaldi and agreed with Clarke's (1970) observation that males achieve larger body size than females. This appears to be an unusual trait for damselfish species (Dulčić and Kraljević, 1995; Kim et al., 2016) and nearshore dioicous reef-associated fish in southern California (Love and Westphal,

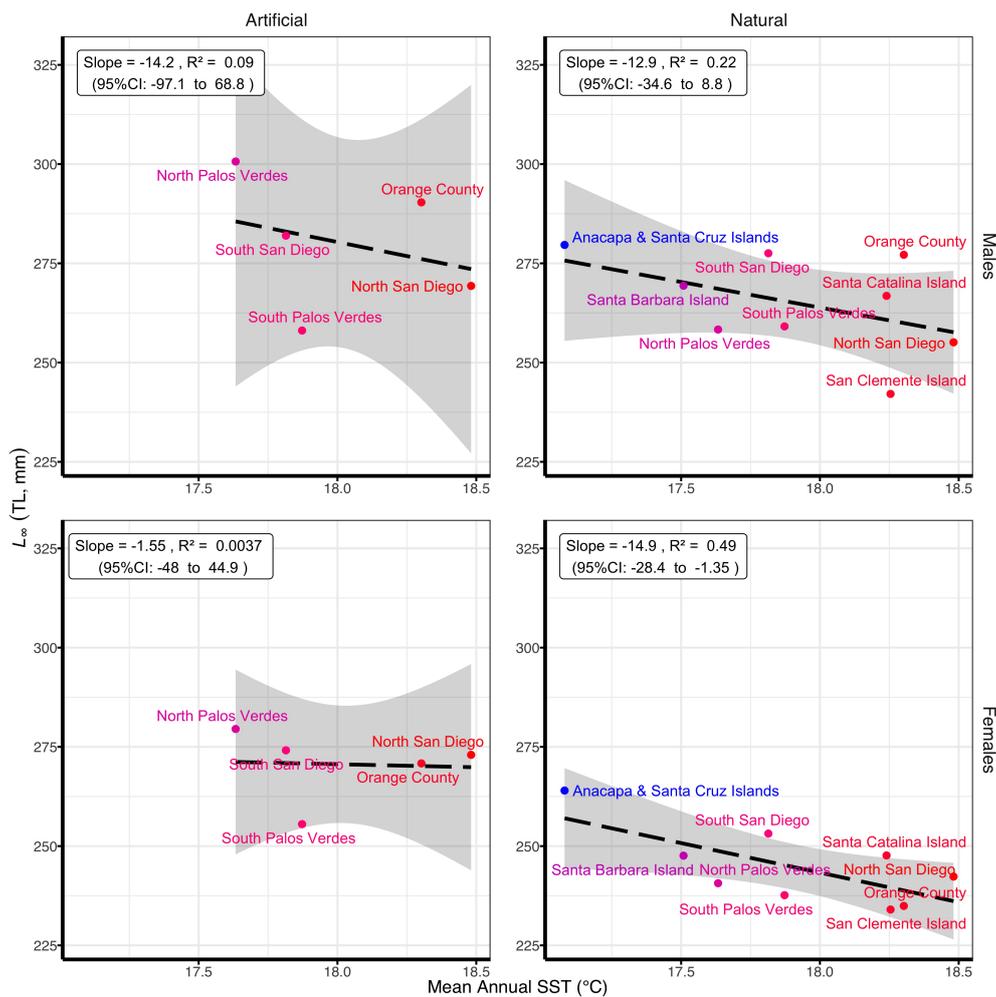


FIGURE 3
Relationship between mean asymptotic total length (L_{∞}) of Garibaldi (*Hypsypops rubicundus*) from the highest ranked (and most complex) model (Table 2) and mean annual sea surface temperature (SST; °C) for January 2002–December 2018. Locations are colored from a scale of blue (cold) to red (warm).

1981; Love et al., 1987; MacNair et al., 2001; Laidig et al., 2003; Pondella et al., 2008; Grebel and Cailliet, 2010; Williams et al., 2012; Miller et al., 2014), where adult females are typically larger. Only one other damselfish species, the Canary Damsel (*Similiparma lurida*) from the northeastern Atlantic (García-Mederos et al., 2016), shows the same pattern of males growing larger than females. Other than average body size, there are no visually discernable phenotypic sexually dimorphic traits for Garibaldi, as researchers were unable to distinguish between sexes both underwater and in the laboratory. In addition, length-weight relationships were very similar between sexes (Figure S4 in Supplementary Material), which demonstrates that males simply grow larger than females, but there are no apparent body shape or weight differences when at similar lengths. Garibaldi exhibit polygyny, where multiple females sequentially lay eggs in one male's nest site. Females visit

multiple nest sites before selecting one for spawning, and tend to prefer nests that have a high percentage of dense turf algae and contain eggs in the early phase of development (i.e., another female had deposited eggs there less than 5 days prior) (Sikkel, 1988; Sikkel, 1989; Sikkel, 1994; Sikkel, 1995). Sikkel 1988; 1989 also found that male size was correlated with female nest preference, and posited that male size is correlated with nest quality, although the relationship between the two variables was not strong. Further, large males may be more successful in the defense of their nesting sites (Sikkel, 1995), suggesting that females are choosing mates based on size.

For mainland locations, Garibaldi grew larger on average (L_{∞}) at artificial reefs than at paired natural reefs. The breakwater and jetty artificial reefs sampled in this study were primarily built to protect coastal infrastructure and were made up of large (1–2 m) quarry rock boulders, providing a complex

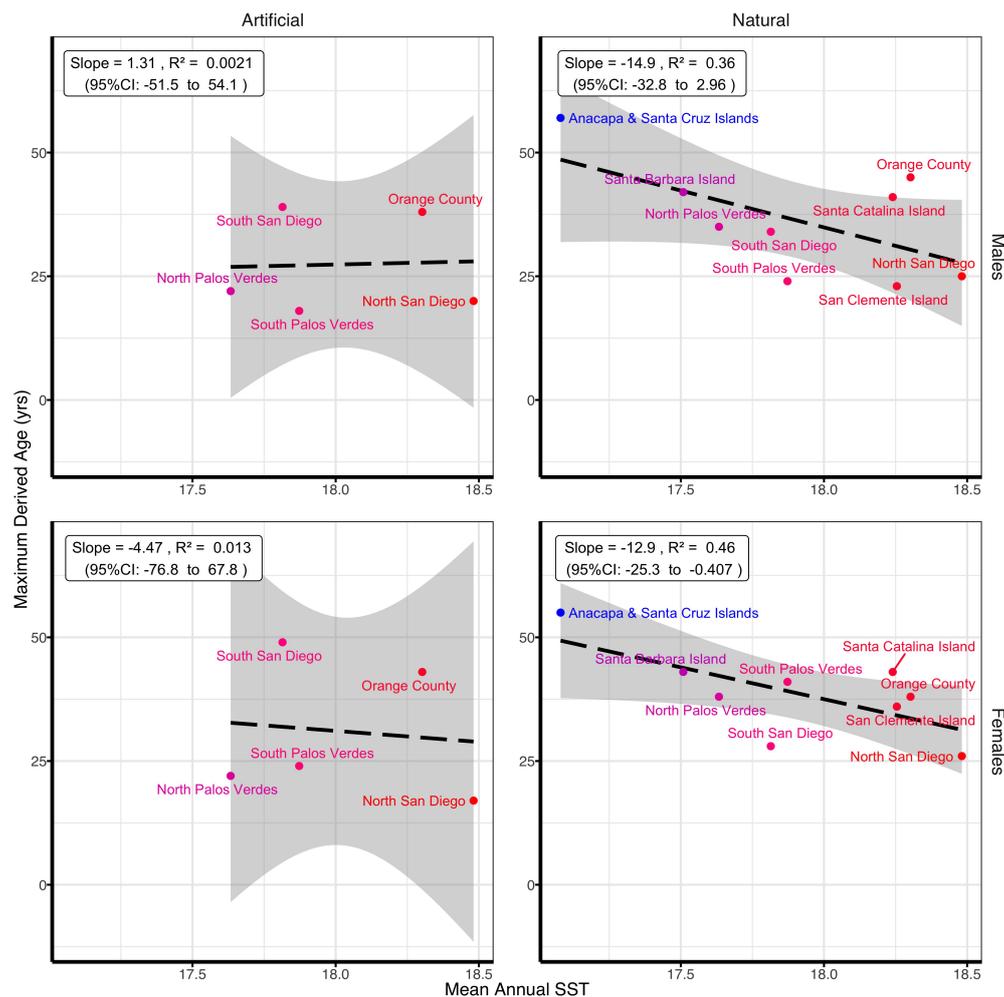


FIGURE 4
Relationship between maximum derived age (life span) of Garibaldi (*Hypsypops rubicundus*) and mean annual sea surface temperature (SST; °C) for January 2002–December 2018. Locations are colored from a scale of blue (cold) to red (warm).

and ideal habitat with vertical rock surfaces and several open spaces for Garibaldi to maintain territories (Stephens et al., 1994; Cook, 2011). Natural rocky reefs within the SCB, although much larger in area, characteristically have lower vertical relief and less habitat complexity than breakwaters (Pondella et al., 2015b). For rocky reefs within the SCB, vertical relief and variation in vertical relief is correlated with increased fish density, biomass, and production (Ambrose and Swarbrick, 1989; Pondella et al., 2002; Pondella et al., 2018; Pondella et al., 2019) indicating that more complex reefs are more productive than homogenous habitats. Garibaldi from artificial reefs grew significantly larger, on average, than fish on natural reefs (sexes pooled, artificial $L_{\infty} = 276$, 95% CI: 271 to 281 mm; natural $L_{\infty} = 252$; 95% CI: 249 to 255 mm) (Table S7 and Figure S10 in Supplementary Material). However, the largest Garibaldi were from natural reefs at Anacapa & Santa Cruz Islands (290 mm TL for females, 307

mm TL for males). This may be due to not sampling artificial habitats in coolest locations, where fish tend to be larger overall, and sampling effort being greater overall for natural reefs ($n = 732$) than for artificial reefs ($n = 298$).

Garibaldi from South Palos Verdes and South San Diego did not exhibit as strong of a difference in growth patterns between natural and artificial habitats as the other mainland locations. Both the natural and artificial reefs at South Palos Verdes are degraded habitats, which may have affected growth. The natural rocky reef (Bunker Point) is characterized by low structural relief and is impacted by sedimentation from ongoing landslides from 1956 to present, resulting in reef burial, scour, and chronic turbidity (Pondella et al., 2018). Scour and sedimentation affect benthic sessile invertebrate communities (e.g., sponges, anemones), the primary component in the diet of Garibaldi, interfering with filter and suspension feeding, or by denuding or

burying the reef and eliminating sessile invertebrates entirely in severe instances (Figurski et al., 2016). In addition, the reduction in light intensity due to persistent turbidity limits benthic algal recruitment and growth (Airoldi and Virgilio, 1998; Gibson and Atkinson, 2003), which is necessary for successful nest-building (Sikkel, 1995). The South Palos Verdes artificial reef (Port of Los Angeles breakwater) borders the busiest harbor in the United States. In South San Diego, both natural (Point Loma) and artificial (Zuniga Jetty) reefs are adjacent to San Diego Bay, a highly urbanized estuary and home to the Port of San Diego. Point Loma is an expansive shelf of natural reef characterized by low relief bedrock with minimal slope and lacking boulders that would provide hiding spaces and good habitat for Garibaldi. Unlike the other breakwaters sampled, Zuniga Jetty is often fully submerged and Garibaldi typically recruit to rocky reefs depths of less than 5m (Clarke, 1970). The distance from the shallow subtidal to the reef where fish were collected at Point Loma, compounded with the relatively high density of adult Garibaldi (Cook, 2011) and skewed sex ratio (2/3 of fish were male) at Zuniga Jetty could explain why natural and artificial reefs in South San Diego are performing similarly in relation to Garibaldi growth patterns.

Growth and longevity patterns followed gradients in mean annual sea surface temperature with fish getting larger and living longer in cooler environments than in warmer environments, particularly for females from natural reefs. This is a common intraspecific pattern in damselfishes and marine fishes in general and has been observed on spatial scales larger and smaller than the present study (Meekan et al., 2001; Choat and Robertson, 2002; Choat et al., 2003; Ruttenberg et al., 2005; Hamilton et al., 2011; Trip et al., 2014; Taylor et al., 2019). Within the SCB, this pattern of fish growing larger at cooler locations has also been observed in the California Sheephead (*Bodianus pulcher*), a fished protogynous hermaphroditic wrasse (Hamilton et al., 2011) and in recreationally and commercially targeted conspecific rockfish species (Scorpaenidae: *Sebastes*) (Love et al., 1990). Further, California Halibut (*Paralichthys californicus*) grew larger in the cooler waters of central California than in southern California (MacNair et al., 2001). The pattern in the present study was not as strong as in other studies, likely due to the mean annual SST values throughout the SCB varying by only $\sim 1.5^{\circ}\text{C}$. Further, sample sizes were relatively small for a given location, especially when grouping populations by sex, which can increase variability in estimates of L_{∞} and T_{max} (Kritzer et al., 2001). The relationship between age and growth and SST was not present for Garibaldi from artificial reefs, but these reefs spanned a lower temperature range than natural reefs. Further, annual and seasonal temperature patterns in the SCB are dynamic (Dong et al., 2009; Kim and Cornuelle, 2015) and long-term annual SST means over the relatively small spatial scale may not be indicative of the environmental conditions that influence Garibaldi life history at local scales. Finally, fish were collected opportunistically over a six-year

period, with some locations only being visited once and others being visited multiple times. Differences in sample dates across some sites could mean that the younger fish in a sample may have experienced different environmental conditions during their early faster growth years prior to when they were collected, possibly confounding temporal and spatial effects on growth. However, this was unlikely to affect the major results of our study. Most natural-artificial reef pairs were collected in the same or similar sets of years. Further, we focused our comparisons mostly on estimates of L_{∞} , and for the long-lived Garibaldi, this parameter is going to primarily be determined by the 10- to 40-year-old fish in our sample which would not be impacted by the relatively small differences in collection dates.

Garibaldi have a much longer lifespan than has previously been reported. The next longest lifespan for a Pomacentrid is for the White-Ear Scalyfin (*Parma microlepis*) endemic to Eastern Australia at 37 years maximum age (Tzioumis and Kingsford, 1999), and most damselfishes considered to be “long-lived” live to be only 10-20 years old (Fowler, 1990; Fowler and Doherty, 1992; Kohda, 1996; Meekan et al., 2001; Schwamborn and Ferreira, 2002; Ruttenberg et al., 2005). Maximum age for both sexes was negatively correlated with mean annual SST for natural reefs, supporting the hypothesis that fish live longer in colder locations. There was also no apparent difference in longevity between sexes although the relationship between mean annual SST and maximum age was stronger for females than for males. The oldest 55-year-old female and 57-year-old male were from the highest latitude and the coolest Santa Cruz Island location. It is important to note that mean annual SST and latitude do not follow the same pattern within the SCB due to the complex current system. The furthest south location, South San Diego had the second oldest female fish (49 years) and is cooler than locations that are further north due to the Southern California Countercurrent (Table 1; Figure 1). There were consistently multiple fish around 40 years old collected from many sites (Table S4, Figure S6 in Supplementary Material). The long lifespan of this species and relatively small sample sizes for each site and sex, make it less likely to collect very old fish and thus limits the ability to make precise longevity comparisons across sites.

This study has demonstrated that habitat and environmental conditions can have a substantial effect on localized age and growth patterns in the absence of fishing effects. Our results contribute to the growing body of evidence that age and growth patterns for marine fish vary spatially and among habitats, even at small geographical scales. In addition, this study provided location-, habitat- and sex-specific parameters enabling researchers to effectively apply appropriate life history traits for specific localities to estimate secondary production (e.g., Claisse et al., 2014; Granneman and Steele, 2014) and as a basis to further investigate differences in habitat quality. The effects of habitat and temperature on both age and growth patterns were stronger for female Garibaldi than for males.

Behavioral factors such as sexual selection and competition amongst males for apparently preferable nesting sites (Sikkel, 1988; Sikkel, 1989; Sikkel, 1995) may play a larger role in shaping the localized life history patterns for male Garibaldi than for females. Additional variability in male growth rates and longevity could result from variability in energy expenditures for territoriality, courting behaviors, nest defense, and/or differences in habitat quality at the scale of individual male territories. That adult female Garibaldi grow significantly larger on artificial reefs is a potential indicator of habitat quality and may have regional population-level implications. Estimates of total reproductive-energy output (i.e., combining fecundity, egg size, egg energy content) per spawn scaled hyperallometrically with body size for Garibaldi and other demersal spawning damselfish species, and therefore larger mothers contribute disproportionately to population replenishment (Barneche et al., 2018). The amount of human-built structures in the marine environment is expected to increase in the coming decades with the continued expansion of marine infrastructure associated with activities such as port construction, marine aquaculture, sea level rise adaptation, and marine renewable energy development (Tickell et al., 2019; Bugnot et al., 2020), and with the growing interest to design artificial reefs for reef habitat restoration and mitigation (e.g., Pondella et al., 2006; Reed et al., 2006; Pondella et al., 2018; Komyakova et al., 2019) and/or for fisheries enhancement (e.g., Polovina and Sakai, 1989; Santos et al., 2011; Roa-Ureta et al., 2019). Fish life history studies that compare ecological performance across habitat types, such as this one, are essential to inform the design and assessment of these future reefing projects.

Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

Ethics statement

The animal study was reviewed and approved by Institutional Care and Use Committees at both Cal Poly Pomona and Occidental College.

Author contributions

CW envisioned the project, managed and contributed to field work and dissections, performed computational analyses, and wrote the manuscript. JE contributed to field work and dissections, prepared histological preparations of gonads and performed microscopic determination of sex. DP contributed to field work and writing. JC envisioned the project, contributed to

field work, and oversaw writing and data analysis. All authors contributed to the article and approved the submitted version.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2022.988158/full#supplementary-material>

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